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Chapter 13

Ambient Level of NO\textsubscript{x} and NO\textsubscript{y} as Indicators of Photochemical Activity in an Urban Center

Alberto Mendoza and Edson R. Carrillo

Additional information is available at the end of the chapter

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1. Introduction

Atmospheric pollution is considered a severe problem, especially in large urban areas where anthropogenic emissions (e.g., emissions from domestic, industrial, and transportation activities, as well as from other productive sectors) mix with biogenic emissions (i.e., emissions with natural origins). Anthropogenic emissions include gas-phase primary air pollutants such as nitric oxide (NO), nitrogen dioxide (NO\textsubscript{2}), carbon monoxide (CO), volatile organic compounds (VOCs), and sulfur dioxide (SO\textsubscript{2}). Although these pollutants can produce harmful health effects, their ability to react as precursors of secondary air pollutants is one of their most relevant characteristics.

Ozone (O\textsubscript{3}) is produced as a secondary pollutant by photochemical reactions occurring between nitrogen oxides (NO\textsubscript{x}, where NO\textsubscript{x} = NO + NO\textsubscript{2}) and VOCs in the presence of sunlight [1, 2]. For a long time, interest in O\textsubscript{3} focused on its direct health effects as a major constituent of photochemical air pollution [3, 4] and its impacts on vegetation [5, 6]. However, O\textsubscript{3} also affects the energy budget of the atmosphere; thus, it has become part of a family of species referred to as short-lived climate pollutants (SLCP) [7].

Photochemical activity occurs in both natural and human-altered environments. Theoretically, in the presence of only NO\textsubscript{x} in the troposphere, O\textsubscript{3} generation could be described by a simple mechanism known as the NO\textsubscript{x}-O\textsubscript{3} photostationary state, summarized by the following reactions:

\[
\text{NO}_2 + h\nu \rightarrow \text{NO} + \text{O}_3
\]
where \( hv \) represents a photon (sunlight energy), \( O^\cdot \) denotes an oxygen free radical (an oxygen atom with an unpaired electron), and \( M \) is known as a third body (molecule) that acts as an energy sink. In the atmosphere, this third body is typically \( N_2 \) or \( O_2 \). Under this chemical reaction scheme, the net \( O_3 \) production is zero, and according to the photostationary state equation, the concentration of \( O_3 \) can be determined based on the concentration of \( NO_x \) and the amount of solar radiation:

\[
\left[ O_3 \right]_{ps} = \left( \frac{j_{NO_2}}{k_{R3}} \right) \left[ NO_2 \right] \left[ NO \right],
\]

where \( k_{R3} \) is the kinetic reaction rate constant for reaction (3) and \( j_{NO_2} \) is the photolysis rate of \( NO_2 \). As a demonstration, if a 10 parts per billion (ppb) value is considered for the term \( j_{NO_2}/k_{R3} \), the \( O_3 \) concentration would be 27 ppb, with an initial value of 100 ppb of NO [8]. An \( O_3 \) concentration ranging between 10–40 ppb is typical of rural areas around the globe [9].

When VOCs are added to the mixture of species present in the troposphere, the observed \( O_3 \) levels are higher than those predicted by the photostationary state formulation. In this case, interactions in the \( O_3 \)-\( NO_x \)-VOC system are initialized by the hydroxyl (HO\(^\cdot\)) radical, through the photolysis of \( O_3 \):

\[
O_3 + hv \rightarrow O_2 + O(1D)
\]

\[
O(1D) + H_2O \rightarrow 2HO^\cdot,
\]

where \( O(1D) \) represents an excited singlet oxygen atom. The oxidation of the VOCs (represented here as \( RH \), where \( R \) is an organic functional group, e.g., an alkyl group) continues in the presence of HO\(^\cdot\) and NO\(_x\) to produce more \( O_3 \) and a variety of nitrogen-containing species. The process can be represented by the following generalized reactions [8]:

\[
RH + HO^\cdot + O_2 \rightarrow RO_2^\cdot + H_2O.
\]

The alkyl peroxide radical (\( RO_2^\cdot \)) reacts with NO to form aldehydes (\( R'CHO \)) and peroxide radicals (\( HO_2^\cdot \)).
Reaction 8 is a relevant process because it represents an alternate route for the production of \( \text{NO}_2 \) from \( \text{NO} \), but without destroying \( \text{O}_3 \) (in contrast to reaction [3]). A similar process occurs when the aldehydes continue reacting to eventually form acyl peroxy radicals (\( \text{R} \cdot \text{C(O)O}_2 \cdot \)) that undergo a similar fate as the \( \text{RO}_2 \cdot \) radicals in reaction (8). Thus, \( \text{O}_3 \) starts to accumulate in the system, as it is no longer destroyed by reaction (3). It continues to be produced by reactions (1) and (2) as \( \text{NO}_2 \) and \( \text{NO} \) cycle through this set of reactions. In addition, the \( \text{RO}_2 \cdot \) radicals can produce nitrates when they react with \( \text{NO}_2 \):

\[
\text{RO}_2 \cdot + \text{NO} + \text{O}_2 \rightarrow \text{NO}_2 + \text{R} \cdot \text{CHO} + \text{HO}_2 \cdot 
\]

(8)

This description is not exhaustive, as we have focused our attention on the main reactions of \( \text{O}_3 \) production by \( \text{NO}_x \) and VOCs. As indicated, \( \text{NO}_x \) acts as catalyst in these reactions, while the VOCs continue to undergo oxidation until they are converted to \( \text{CO}_2 \). In parallel, a variety of different inorganic (e.g., \( \text{HNO}_3 \)) and organic (e.g., \( \text{RONO}_2 \)) nitrogen-containing species are also produced. Some of these substances act as reservoirs of \( \text{NO}_x \) which are released to the reacting mixture upon decomposition (e.g., peroxy acetyl nitrate or PAN, an organic nitrate), and others act as sinks (most notably, \( \text{HNO}_3 \)). The sum of \( \text{NO}_x \) and these additional inorganic and organic nitrates is referred to as total reactive nitrogen or total odd nitrogen oxides (\( \text{NO}_y \)).

In general, the \( \text{O}_3 \)-\( \text{NO}_x \)-VOC system increases the \( \text{NO}_x \) available to react through reactions (1)-(2) (increasing \( \text{O}_3 \)), thereby producing a complex mixture of partially oxidized VOCs that mixes with freshly emitted VOCs and different oxidized nitrogen species (\( \text{HNO}_3, \text{HNO}_2, \text{NO}_3, \text{PAN}, \) etc.). The study of the dynamics of \( \text{O}_3 \) production can be quite challenging, given the complexity of the chemical mixtures and their nonlinear response to emission changes and meteorological conditions. For this reason, comprehensive air quality models have been devised to study the complex physical and chemical processes that participate in gas-phase production of \( \text{O}_3 \) and other air pollutants [10].

Observational-based approaches have proven to be valuable for describing the conditions and regimes that foster air pollution. Some observational approaches use considerable amounts of data that can be obtained through networks of routine air quality and meteorological monitoring stations, and conclusions are then inferred based on the statistical analysis of these data [11-17]. The analyses of these databases can be enhanced if information is also available for additional indicator species that are typically not routinely monitored, such as \( \text{NO}_y \).

In this study, we have analyzed data gathered by the routine air quality monitoring stations of the Monterrey Metropolitan Area (MMA; 25° 40’ N, 100° 18’ W). Monterrey is the third most populated urban center in Mexico (4.1 million inhabitants), second in size in the country in...
terms of industrial infrastructure, and one of the cities with the worst air quality problems in Mexico [18]. The MMA (see Figure 1) has been in violation of the 1-hr Mexican Air Quality Standard for $O_3$ (0.11 ppm) since the establishment of its routine air quality monitoring system in 1993. For example, the 1-hr $O_3$ standard was exceeded on 48 different days in 2011 [19]. Peak $O_3$ concentrations can reach 170 ppb and typically occur at the downtown or western air quality stations [20]. Some studies report that the MMA is the fifth most polluted Latin American city in terms of $O_3$ [21]. In addition to the large number of industrial facilities located inside or nearby the metropolitan area that contribute to the poor air quality, emissions from several gas-fired electric utilities and one of the six refineries that operate in the country (located less than 40 km to the east of downtown Monterrey) add to the anthropogenic burden imposed on the airshed. Despite the importance of the contribution of area and point sources to the total emissions inventory, the mobile sources represent approximately 75% of the total anthropogenic emissions released in the MMA [20]. This is a result of a relatively high proportion of vehicles per inhabitant registered in this urban center (approximately one vehicle for every two inhabitants).

2. Methods: Database and analysis tools

The study used data collected by the Sistema Integral de Monitoreo Ambiental (SIMA; Integrated Environmental Monitoring System) of the MMA. At the time of the study, valid data from six operational routine air quality stations were available: Downtown, Southeast, Southwest, Northeast, Northwest, and North (see Figure 1). Data archived for the period of August 2012 to August 2013 were retrieved. The database contained hourly-average concentrations of CO, NO, NO$_2$, and $O_3$, as well as meteorological parameters (relative humidity, atmospheric pressure, dry bulb temperature, solar radiation [SR], wind speed [WS], and wind direction [WD]). These data go through a quality assurance/quality control process defined by SIMA in compliance with international standards. Additional air quality parameters monitored by the stations (mainly, SO$_2$ and particulate matter with aerodynamic diameter less than or equal to 10 microns [PM$_{10}$] and less than or equal to 2.5 microns [PM$_{2.5}$]) were not used in this study, as the focus was on the relationship between $O_3$ and NO$_x$.

In addition to the above chemical parameters, NO$_y$ was monitored in the Downtown station from August 2012 to August 2013, using a NO-NO$_y$ (Thermo Scientific, Model 42i-Y) chemiluminescence continuous sampling device. NO$_y$ is a chemical parameter that is not routinely measured by Mexican air quality stations, as it is not considered a criteria pollutant. The decision to deploy the NO-NO$_y$ instrument at the location of the Downtown station was based on spatial homogeneity studies performed in the past for the MMA [22], which indicate that pollutant levels observed in the Downtown station are representative of the MMA, with the exception of the Southeast region. The data collected by the NO-NO$_y$ device went through a validation process similar to that routinely conducted by SIMA on its own collected data. The main conditions to reject data included the following red flags: obstructed capillary tubing, low flow in the inlet, and concentrations outside the measurement range (i.e., above 200 ppbv). The data were then consolidated and analyzed on a daily basis, a monthly basis, and by seasons (Summer...
and Fall 2012, and Winter, Spring, and Summer 2013). Summer 2012 included only data from August and September of 2012; Summer 2013 did not include data for September 2013.

Figure 1. Municipalities that comprise the Monterrey Metropolitan Area (left panel) and location of the air quality stations monitoring stations used in this study (right panel): 1 Southeast, 2 Northeast, 3 Downtown, 4 Northwest, 5 Southwest; data from the North station was not used

Once the data were consolidated, an exploratory analysis of the data set was conducted to derive descriptive statistics. In addition, correlation analysis was conducted to explore relations between the chemical and meteorological parameters, with emphasis on the relationships among O$_3$, NO$_x$, and NO$_y$. This analysis was complemented by the use of polar plots and wind roses to determine the relationship between high pollutant levels and transport conditions (wind speed and direction).

Grouping techniques were then applied for further exploration of the data. Two methods were used: Principal Components Analysis (PCA) and Analysis of Variance (ANOVA). PCA is helpful in reducing the dimensionality of the dataset, and ANOVA can identify differences among data groups. Thus, PCA was used to identify the most important variables in the dataset which in turn merit further exploration. With the ANOVA, we attempted to determine differences in the weekday and weekend conditions that resulted in high O$_3$ levels. All statistical analyses were conducted using Minitab$^\text{®}$ 16, and plots were constructed using the R programming language.

3. Results

3.1. Descriptive statistics

Table 1 summarizes the descriptive statistics of the chemical and meteorological parameters retrieved from the Downtown air quality monitoring station. The values presented are based
on the daily means for the August 2012-August 2013 period. The climatology of the MMA is portrayed in these results: a hot semi-arid region characterized by extreme weather conditions [23]. The long-term average temperature of Monterrey is around 23°C, with temperatures that can reach 40°C in summer and below 0°C in winter. Average humidity is 62%; the rainy season is between August and October, directly related to the occurrences of hurricanes in the Gulf of Mexico (the average mean annual precipitation is around 600 mm). Due to its geographic location, the MMA is influenced by anti-cyclonic systems from the Gulf of Mexico [24], which in some instances result in high atmospheric stability in the region, thereby inhibiting the vertical mixing of pollutants.

<table>
<thead>
<tr>
<th>Species/Parameters</th>
<th>Mean ± Std. Dev.</th>
<th>Median</th>
<th>Maximum</th>
<th>Minimum</th>
</tr>
</thead>
<tbody>
<tr>
<td>O₃ (ppb)</td>
<td>25.7 ± 9.0</td>
<td>24.4</td>
<td>50.7</td>
<td>2.6</td>
</tr>
<tr>
<td>NO₂ (ppb)</td>
<td>16.3 ± 5.9</td>
<td>15.3</td>
<td>39.2</td>
<td>3.3</td>
</tr>
<tr>
<td>NO (ppb)</td>
<td>6.5 ± 4.7</td>
<td>5.3</td>
<td>34.4</td>
<td>0.6</td>
</tr>
<tr>
<td>NOₓ (ppb)</td>
<td>40.0 ± 20.3</td>
<td>35.2</td>
<td>188.2</td>
<td>12.1</td>
</tr>
<tr>
<td>CO (ppm)</td>
<td>0.78 ± 0.26</td>
<td>0.75</td>
<td>1.78</td>
<td>0.21</td>
</tr>
<tr>
<td>Solar radiation (kW/m²)</td>
<td>0.24 ± 0.12</td>
<td>0.24</td>
<td>0.70</td>
<td>0.00</td>
</tr>
<tr>
<td>Temperature (°C)</td>
<td>24.4 ± 5.2</td>
<td>25.8</td>
<td>32.5</td>
<td>9.5</td>
</tr>
<tr>
<td>Wind speed (km/h)</td>
<td>1.84 ± 0.47</td>
<td>1.84</td>
<td>3.21</td>
<td>0.71</td>
</tr>
<tr>
<td>Relative Humidity (%)</td>
<td>62.9 ± 16.8</td>
<td>64.17</td>
<td>98.9</td>
<td>13.5</td>
</tr>
</tbody>
</table>

Table 1. Descriptive statistics for parameters monitored at the Downtown air quality station of the MMA (annual means, medians, maximum values, and minimum values of the daily averages)

Details of the monthly variation of the daily averages of the chemical and meteorological parameters are depicted in Figure 2. O₃ levels tend to be the highest during springtime, as well as during late summer and early fall. These periods correspond to the times when temperature and SR are high and wet precipitation is low, i.e., conditions that foster photochemical reactions. Another relevant feature presented in Figure 2 involves the inverse relationship observed between O₃ and NOₓ. That is, during the cold months, when SR is low, NOₓ levels tend to be high due to inhibition of the reactions between VOCs and NOₓ that promote the conversion of NO to NO₂, as well as suppression of the photo-dissociation of NO₂ to produce O₃. We need to bear in mind that NOₓ emissions are mainly in the form of NO. Thus, an accumulation of NO and NO₂ is observed in winter. At other times, temperatures and SR increase, promoting photochemical reactions. A meteorological component influences the observed NOₓ concentrations: shallow mixing heights and low wind speeds during winter. The O₃ seasonal boxplots indicate that even though fall is the second most important season with respect to the frequency of peak O₃ levels, it is also the time when the average O₃ is the lowest (Figure 3). This can be explained by the incidence of rain events associated with this season.
Figure 2. Time series plots of monthly averages of the mean daily values of chemical and meteorological parameters observed at the Downtown station.
3.2. Spatial analysis

The exploratory analysis was complemented by the construction of bivariate polar plots of NO\textsubscript{x} (see Figure 4) and wind roses (see Figure 5) for five of the six air quality monitoring stations, covering the main sub-regions of the MMA. The five selected stations correspond to typical upwind locations (Northeast and Southeast), Downtown Monterrey, and typical downwind locations (Northwest and Southwest), as illustrated in Figure 5. Seasonal variation is also presented in these plots.

The polar plots represent NO\textsubscript{x} concentration as a function of wind speed and direction, and they help identify the possible occurrence/prevalence of horizontal transport conditions that lead to high NO\textsubscript{x} events. Further away from the center of the plot, the wind speed is higher. The form of the polar plot can provide an indication of the wind direction and speed frequency; however, this information is better observed through the corresponding wind roses.

![Figure 3. O\textsubscript{3} boxplots by seasonal periods for 1-hr average values reported at the Downtown station](image)

In general, Figure 4 indicates higher transport of NO\textsubscript{x} from the northern stations (this can be catalogued as “fresh” NO\textsubscript{x}), which then is consumed as the air masses traverse the urban core. The transport patterns appear to be similar in all seasons. Thus, the change in NO\textsubscript{x} by season appears to be more strongly linked to photochemical conditions (temperature and SR) than to wind patterns.

Using the same analysis as was used for NO\textsubscript{x}, Figure 6 depicts the behavior of O\textsubscript{3} as a function of wind speed and direction exclusively at the Downtown station. In all cases, the higher concentrations begin to occur when the wind velocity was above 2 m/s, indicating the influence of transport and aging of air masses on the observed O\textsubscript{3} concentrations and weak O\textsubscript{3} production.
under stagnant conditions. Transport influences the levels of O$_3$; consequently, the highest O$_3$ levels correlate with the prevailing wind directions (wind blowing from the Northeast-East-Southeast and from the North-Northwest-West). The East-to-West air flow channels air masses through the long axis of the MMA, adding more precursors to the chemically aged mixtures as they traverse the urban core, thereby promoting O$_3$ production.

The frequency of high O$_3$ concentrations related to a specific wind direction cannot be inferred from the polar plots, however. That analysis was conducted by constructing boxplots for 1-hr average O$_3$ concentrations for the three main wind categories (classified based on direction from which the wind was blowing) and segregated by season (see Figure 7). The year around the peak O$_3$ concentrations and the maximum 1-hr average concentrations occur when the wind blows from the Northeast-East-Southeast. O$_3$ levels, on average, are lower when the wind blows from the North-Northwest-West, and they are the lowest when winds blow from the South-Southeast (with the exception of the summer, when this direction provides higher average concentrations than the North-Northwest-West direction).

Comparison of the O$_3$ levels presented by the polar plots (Figure 6) for spring and fall indicates that the influence of transport is less important during fall than in spring, as wind speeds tend to be lower; high O$_3$ events are mainly related to events when the speed is in the range of 2.0-3.5 m/s. In spring, O$_3$ levels can remain high at wind speeds above 4 m/s.

![Figure 4. Polar plot for 1-h average NO$_x$ concentrations at five different monitoring stations within the MMA (the radial dimension is an indication of increasing wind speed [m/s])](image-url)
Figure 5. Wind roses at five different monitoring stations within the MMA.
3.3. Regression analysis

Figure 2 gives an initial indication of a direct relationship between NO and NO₂ levels and an inverse relationship between NOₓ and O₃ levels. This was explored further by means of scatter plots (not presented for brevity), which were used to derive linear regression expressions among NOₓ constituents and between NOₓ and O₃:

\[ \left[ NO_x \right] = 1.5 \left[ NO \right] + 13.0 \]  
\[ \left[ O_3 \right] = -0.37 \left[ NO_x \right] + 34 \]

where \( i \) represents the molar concentration of species \( i \). Here, the 1-hr average concentrations were used to derive equations (11) and (12). These expressions corroborate prior findings. The relevance of equation (12) is that it provides a measure of the expected, first order change in O₃ provided by a change in NOₓ concentration (which can occur from a change in NO emissions). The results indicate that O₃ would, in fact, increase as NOₓ tends to be reduced. This counterintuitive finding is a well-known result seen in urban areas dominated by a VOC-sensitive regime [25]. Other studies that have used complex photochemical air quality models have reached similar conclusions. Downtown Monterrey appears to be dominated by a VOC-sensitive regime in which VOC controls would work better for reducing O₃ levels [20].
3.4. Analysis of NO\textsubscript{y} levels

Up to this point, our analysis has focused on describing the relationship between NO\textsubscript{x} and \( O_3 \) production based on the meteorological conditions that foster photochemical reactions. The generalized chemical mechanism that describes this relationship indicates that other reaction products can be of interest in describing the dynamics of \( O_3 \) production. In particular, the levels of NO\textsubscript{y} provide additional information and serve as an indicator of the degree to which the air masses have been subjected to photochemical processing (aging), thereby providing information on \( O_3 \) production.

Table 2 presents the results of the correlation analysis of different chemical and meteorological parameters with NO\textsubscript{y}. A correlation coefficient (\( R^2 \)) of NO\textsubscript{2} with NO\textsubscript{y} above 0.78 is noted for months in which \( O_3 \) production is not the highest, and between 0.40 and 0.68 for the months with the highest \( O_3 \) production. In general, a high percentage of NO\textsubscript{y} corresponds to NO\textsubscript{x} (Figure 2); thus, the response of one clearly depends on the other.

The difference between NO\textsubscript{y} and NO\textsubscript{x} (known as NO\textsubscript{z}) includes the inorganic and organic nitrates derived from photochemical processing of the NO\textsubscript{x}-VOCs mixtures. The typical composition of NO\textsubscript{z} is approximately 50%-55% HNO\textsubscript{3}, 35%-40% PAN, 1%-5% HNO\textsubscript{2}, and a marginal contribution of other nitrates [26, 27]. During the cold months, the production of NO\textsubscript{z} can be inhibited by meteorological conditions that do not foster photochemical reactions. Thus, the relative contribution of NO\textsubscript{z} to NO\textsubscript{y} is the highest, and, therefore, the responses of both species are expected to show high correlation. When conditions favor \( O_3 \) production, more NO\textsubscript{2} is produced, thereby limiting the contribution of NO\textsubscript{z} to the total NO\textsubscript{y}. Consequently, the level of correlation drops as the nonlinear photochemistry and sinks of NO\textsubscript{y} increase in importance. Of interest is the fact that during spring, the season with the highest peak \( O_3 \) 1-hr concentrations, the correlation of NO\textsubscript{z} species with NO\textsubscript{y} is the lowest. A similar response is obtained between CO and NO\textsubscript{y}, which can be explained by the typical strong correlation.
between CO and NO\textsubscript{x} emissions, particularly in urban areas that are heavily influenced by mobile-source emissions. With respect to meteorological variables, the highest correlation was with wind speed, and, as in the case of NO\textsubscript{x} and CO, the correlation was the lowest during spring.

<table>
<thead>
<tr>
<th></th>
<th>Summer 2012</th>
<th>Fall 2012</th>
<th>Winter 2013</th>
<th>Spring 2013</th>
<th>Summer 2013</th>
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<tr>
<td>NO\textsubscript{y} (ppbv)</td>
<td>0.811</td>
<td>0.676</td>
<td>0.819</td>
<td>0.411</td>
<td>0.783</td>
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<tr>
<td>NO (ppbv)</td>
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<td>0.661</td>
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<tr>
<td>O\textsubscript{3} (ppbv)</td>
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<td>0.106</td>
<td>0.151</td>
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<tr>
<td>CO (ppmv)</td>
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<td>0.664</td>
<td>0.227</td>
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<td>(&lt;0.001)</td>
<td>(&lt;0.001)</td>
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</tr>
<tr>
<td>Solar radiation (kW/m\textsuperscript{2})</td>
<td>0.011</td>
<td>0.001</td>
<td>0.004</td>
<td>0.001</td>
<td>0.009</td>
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<td>(&lt;0.001)</td>
<td>(0.639)</td>
<td>(0.052)</td>
<td>(0.165)</td>
<td>(&lt;0.001)</td>
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<td>Temperature (°C)</td>
<td>0.009</td>
<td>0.024</td>
<td>0.022</td>
<td>0.005</td>
<td>0.063</td>
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<tr>
<td></td>
<td>(0.048)</td>
<td>(&lt;0.001)</td>
<td>(&lt;0.001)</td>
<td>(0.002)</td>
<td>(&lt;0.001)</td>
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<tr>
<td>Wind speed (km/h)</td>
<td>0.252</td>
<td>0.247</td>
<td>0.203</td>
<td>0.112</td>
<td>0.225</td>
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</tbody>
</table>

Table 2. Correlation coefficients ($R^2$) of parameters monitored by the SIMA downtown station with NO\textsubscript{y} (in parenthesis $p$-values at an $\alpha$ of 0.05)

The results of the correlation analysis suggest a relationship between NO\textsubscript{y} and NO\textsubscript{x}, but a correlation between NO\textsubscript{y} and O\textsubscript{3} is less evident. Figure 2 shows two peaks in the annual cycle of the NO\textsubscript{y} monthly average concentration (constructed from daily NO\textsubscript{y} average concentrations). One peak occurs during the cold months and is associated with the high levels of NO\textsubscript{x} (low NO\textsubscript{z} component), while the second peak appears during spring, which is the period with highest photochemical activity (highest O\textsubscript{3} levels); thus, as NO\textsubscript{x} is consumed, and its levels decrease and NO\textsubscript{y} levels are high, NO\textsubscript{z} must be contributing heavily to NO\textsubscript{y}. The peaks of NO\textsubscript{y} are associated with different conditions, so we looked in greater detail into the seasonal levels of NO\textsubscript{y}. Figure 8 depicts the 1-hr NO\textsubscript{y} concentrations by season. The highest levels (both of peak and season-average concentrations) occur in winter, followed by fall and spring (when O\textsubscript{3} is the highest), with summer registering the lowest levels. The O\textsubscript{3} levels presented in Figure 3 do not appear to track the NO\textsubscript{y} levels.

We then attempted to determine a clearer relationship between O\textsubscript{3} and NO\textsubscript{y} by conducting a Principal Components Analysis (PCA) (see Table 3). The first principal component (PC1) has to do with O\textsubscript{3} precursors in the form of primary emissions or intermediate products (NO\textsubscript{x}, NO\textsubscript{y}, and CO), in line with the results from the descriptive statistics. The second principal component (PC2) covers photochemical O\textsubscript{3} production (O\textsubscript{3} and solar radiation). The third principal component (PC3) refers to a combined meteorological effect that includes horizontal
transport (wind direction) and temperature— an effect that was also observed in the descriptive statistics results. The first two principal components represent almost 60% of the variance.

<table>
<thead>
<tr>
<th>Component</th>
<th>PC1</th>
<th>PC2</th>
<th>PC3</th>
</tr>
</thead>
<tbody>
<tr>
<td>NO\textsubscript{y}</td>
<td>0.481</td>
<td>0.055</td>
<td>-0.009</td>
</tr>
<tr>
<td>NO\textsubscript{2}</td>
<td>0.387</td>
<td>0.087</td>
<td>-0.123</td>
</tr>
<tr>
<td>NO</td>
<td>0.375</td>
<td>0.029</td>
<td>0.151</td>
</tr>
<tr>
<td>O\textsubscript{3}</td>
<td>-0.226</td>
<td>0.447</td>
<td>0.075</td>
</tr>
<tr>
<td>CO</td>
<td>0.400</td>
<td>0.110</td>
<td>0.081</td>
</tr>
<tr>
<td>Solar radiation</td>
<td>-0.090</td>
<td>0.497</td>
<td>0.031</td>
</tr>
<tr>
<td>Temperature</td>
<td>-0.197</td>
<td>0.370</td>
<td>0.411</td>
</tr>
<tr>
<td>Wind velocity</td>
<td>-0.288</td>
<td>0.276</td>
<td>0.084</td>
</tr>
<tr>
<td>Wind speed</td>
<td>0.175</td>
<td>0.339</td>
<td>0.794</td>
</tr>
<tr>
<td>Accumulated variance (%)</td>
<td>0.395</td>
<td>0.595</td>
<td>0.672</td>
</tr>
</tbody>
</table>

Given these results, we conducted an additional assessment on the effects of meteorological parameters on O\textsubscript{3} and NO\textsubscript{y} levels, which showed the effects of horizontal transport. O\textsubscript{3}, temperature, wind speed, and wind direction data were particularly correlated (see Table 4). In addition, complementary polar plots were constructed for NO\textsubscript{y} (Figure 9). The highest correlation between O\textsubscript{3} and wind speed occurred during the winter of 2013. In this case, the
slope of the linear regression was positive (i.e., as the wind speed increases, O\textsubscript{3} also increases). In a similar way, the highest correlation between O\textsubscript{3} and wind direction occurred during the winter of 2013. This indicates that transport is a major factor responsible for the average O\textsubscript{3} levels observed in the winter in the MMA. The polar plot for NO\textsubscript{y} confirms that high levels of NO\textsubscript{y} during winter (which are, as discussed, mostly NO\textsubscript{x}) are associated with low wind speeds. This relationship between NO\textsubscript{y} and wind speed would indicate that NO\textsubscript{x} is not being processed and translated to high local production of O\textsubscript{3} in the winter.

An additional way to establish the influence of transport on the levels of air pollutants is to analyze the CO-to-NO\textsubscript{y} ratio. Values close to 10 are indicative of an influence of local sources, and ratios larger than 100 indicate the influence of regional (remote) sources [11]. Figure 10 depicts the CO-to-NO\textsubscript{y} ratios observed at the Downtown station as a function of the season. The average ratio tends to be the inverse with respect to the behavior of the average levels of NO\textsubscript{y}, although one main difference exists: the lowest ratios are observed during winter but also during spring, intermediate values are observed during the fall, and the highest average values are registered during the summer. In a few cases, the CO-to-NO\textsubscript{y} ratio reached values of 100 or more, and the 95\textsuperscript{th} percentile, in general, was below 56. Thus, air pollution in the MMA during winter appears to be heavily influenced by low wind conditions (Figure 2 indicates the lowest average wind speeds during this season), and poor photochemical activity limits the levels of NO\textsubscript{x} and O\textsubscript{3}. The O\textsubscript{3} levels that are registered are characteristic of background amounts.

<table>
<thead>
<tr>
<th>Wind speed</th>
<th>Summer 2012</th>
<th>Fall 2012</th>
<th>Winter 2013</th>
<th>Spring 2013</th>
<th>Summer 2013</th>
</tr>
</thead>
<tbody>
<tr>
<td>R\textsuperscript{2}</td>
<td>0.212</td>
<td>0.271</td>
<td>0.350</td>
<td>0.025</td>
<td>0.231</td>
</tr>
<tr>
<td>Slope</td>
<td>2.348</td>
<td>3.625</td>
<td>3.342</td>
<td>1.164</td>
<td>2.986</td>
</tr>
<tr>
<td>Intercept</td>
<td>6.836</td>
<td>0.937</td>
<td>3.319</td>
<td>35.560</td>
<td>4.850</td>
</tr>
<tr>
<td>p-value</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
</tr>
</tbody>
</table>

<table>
<thead>
<tr>
<th>Wind direction</th>
<th>Summer 2012</th>
<th>Fall 2012</th>
<th>Winter 2013</th>
<th>Spring 2013</th>
<th>Summer 2013</th>
</tr>
</thead>
<tbody>
<tr>
<td>R\textsuperscript{2}</td>
<td>0.001</td>
<td>0.022</td>
<td>0.216</td>
<td>0.002</td>
<td>0.010</td>
</tr>
<tr>
<td>Slope</td>
<td>-0.005</td>
<td>-0.031</td>
<td>-0.088</td>
<td>0.014</td>
<td>-0.027</td>
</tr>
<tr>
<td>Intercept</td>
<td>25.432</td>
<td>24.893</td>
<td>35.313</td>
<td>44.662</td>
<td>28.428</td>
</tr>
<tr>
<td>p-value</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
</tr>
</tbody>
</table>

<table>
<thead>
<tr>
<th>Temperature</th>
<th>Summer 2012</th>
<th>Fall 2012</th>
<th>Winter 2013</th>
<th>Spring 2013</th>
<th>Summer 2013</th>
</tr>
</thead>
<tbody>
<tr>
<td>R\textsuperscript{2}</td>
<td>0.454</td>
<td>0.327</td>
<td>0.341</td>
<td>0.344</td>
<td>0.536</td>
</tr>
<tr>
<td>Slope</td>
<td>2.834</td>
<td>1.877</td>
<td>2.017</td>
<td>2.031</td>
<td>3.792</td>
</tr>
<tr>
<td>p-value</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
<td>&lt; 0.001</td>
</tr>
</tbody>
</table>

Table 4. O\textsubscript{3} correlations with meteorological parameters (p-values at an \( \alpha \) of 0.05)
that are transported to the urban center when wind speeds are sufficiently strong. The spring level is also heavily influenced by local emissions, but meteorological conditions do favor photochemical activity, resulting in production of the highest levels of O$_3$. Fall would seem to be influenced partly by local emissions and partly by transport. Finally, summer is characterized by the influence of more regional (and aged) emissions. During the first two months of the summer season, the ambient conditions tend to be characterized by high temperatures and deep planetary boundary layers that promote vertical mixing. In addition, conditions favor the mixing of these vertical columns with pollutants that undergo long-range transport.

**Figure 9.** Polar plot for NO$_x$ at the Downtown monitoring station

**Figure 10.** Seasonal variation of CO-to-NO$_x$ ratios (the red line represents a value of 10 for the ratio)
3.5. Weekend-weekday effect

A final aspect of interest in this study was to define if a weekend-weekday effect could be established based on NO\textsubscript{x} and NO\textsubscript{y} readings. Previous studies have demonstrated that O\textsubscript{3} levels in urban centers can be higher during weekends [28, 29], a condition that might be thought of as counterintuitive, as emissions from anthropogenic sources tend to be lower during weekends. An explanation for the occurrence of this weekend-weekday effect can be obtained by analyzing the photochemical mechanism for the production of O\textsubscript{3}. Urban centers tend to be VOC-sensitive; that is, NO\textsubscript{x} is in excess and, thus, processes that remove NO\textsubscript{2} from the reactive mixtures are promoted. For example, NO\textsubscript{2} efficiently scavenges HO\textsuperscript{*} radicals through reaction 9, producing HNO\textsubscript{3} which is a direct sink for NO\textsubscript{x}. This limits the reaction of HO\textsuperscript{*} radicals with VOCs, which, in turn, controls the production pathways for O\textsubscript{3}. During weekends, the ratio at which VOCs and NO\textsubscript{x} are emitted can change, preferentially reducing the emissions of NO\textsubscript{x}. If this happens, less NO\textsubscript{x} is available in the atmosphere, reaction 9 becomes less important, more HO\textsuperscript{*} radicals become available to react with the VOCs, and, thus, O\textsubscript{3} tends to increase. Figure 11 illustrates the distribution of NO\textsubscript{y} concentrations by day of the week. At a first glance, NO\textsubscript{y} values tend to be quite constant throughout the day. However, Tuesdays and Wednesdays present the lowest peak values, while average levels on Sunday tend to be lower than during the rest of the week. If the average NO\textsubscript{x}-to-NO\textsubscript{y} ratio is examined, values tend to be lower during the weekends (Figure 12). The NO\textsubscript{x}-to-NO\textsubscript{y} ratio provides information on the level of photochemical processing of the observed air parcels. Ratios closer to 1 would indicate that most of the NO\textsubscript{y} is in the form of NO\textsubscript{x} and thus, it could imply the presence of fresh emissions or the transport of unprocessed emissions. As the ratio drops, the relative contribution of NO\textsubscript{y} increases, indicating that an aged (photochemically processed) air parcel is being observed.

![Figure 11. Boxplot for NO\textsubscript{y} levels by day of the week](http://dx.doi.org/10.5772/59752)
Finally, we conducted an Analysis of Variance (ANOVA) test for the NO$_x$-to-NO$_y$ ratio by season (see Table 5). For the hot months (spring and summer), lower ratios were observed during weekends compared to weekdays. In fact, the lowest average ratio was obtained during spring, the season with the highest average and peak O$_3$ concentrations. In contrast, the winter of 2013 showed a marginally higher ratio during the weekends than on the weekdays, but the difference was not statistically significant. This could be the result of an intensive use of fuels during the whole week in response to the low temperatures. For the fall of 2012, this statistical test also indicates no difference between weekdays and weekends. Fall, as the results from previous sections suggest, is influenced by days with high O$_3$ production and days with very low O$_3$ levels as a result of the transition from a period with high SR and temperature to the regional rainy season.

<table>
<thead>
<tr>
<th>Season</th>
<th>Average ± Standard Dev. (Weekdays)</th>
<th>Average ± Standard Dev. (Weekends)</th>
<th>p-value</th>
<th>Difference between Weekdays and Weekends</th>
</tr>
</thead>
<tbody>
<tr>
<td>Summer 2012</td>
<td>0.606 ± 0.139</td>
<td>0.582 ± 0.124</td>
<td>0.033</td>
<td>Yes</td>
</tr>
<tr>
<td>Fall 2012</td>
<td>0.671 ± 0.141</td>
<td>0.668 ± 0.107</td>
<td>0.657</td>
<td>No</td>
</tr>
<tr>
<td>Winter 2013</td>
<td>0.545 ± 0.086</td>
<td>0.549 ± 0.077</td>
<td>0.504</td>
<td>No</td>
</tr>
<tr>
<td>Spring 2013</td>
<td>0.533 ± 0.112</td>
<td>0.496 ± 0.142</td>
<td>&lt; 0.001</td>
<td>Yes</td>
</tr>
<tr>
<td>Summer 2013</td>
<td>0.586 ± 0.110</td>
<td>0.561 ± 0.112</td>
<td>&lt; 0.001</td>
<td>Yes</td>
</tr>
</tbody>
</table>

Table 5. Results of the ANOVA test for the NO$_x$-to-NO$_y$ ratio (p-values are at an α of 0.05)
4. Conclusions

Observational-based methods have proven valuable for analysis of the chemical interactions that give rise to high air pollution events in urban areas. In particular, the vast amount of information gathered by networks of air quality stations can provide insight into the production of secondary air pollutants such as $\text{O}_3$, as atmospheric conditions change throughout the year. This analysis can be enhanced if complementary observations are also made of species that are not routinely registered, such as $\text{NO}_y$.

In this study, we presented a statistical analysis performed on the air quality and meteorological data registered by the routine air quality stations of the MMA. The analysis included descriptive statistics, regression analysis, correlation analysis, PCA, and ANOVA, along with the interpretation of bivariate polar plots, wind roses, and boxplots. In addition, ratios of $\text{NO}_x$ with $\text{NO}_y$ and CO with $\text{NO}_y$ provided additional information on the level of chemical processing of the air masses traversing the MMA. When used together, these techniques prove to be complementary, thus providing more robust results.

In the MMA, $\text{O}_3$ registers two distinct annual peaks: one in spring and the other in late summer-early fall. The analysis of meteorological conditions and air pollutant levels indicate that the $\text{O}_3$ concentrations during winter would be characteristic of background conditions that get transported to the urban center when wind speeds are sufficiently strong. $\text{O}_3$ production in winter is small because typical conditions that foster photochemical activity are not present, such as relatively high temperatures and strong solar radiation. During this season, $\text{NO}_y$ is composed mainly of locally-emitted $\text{NO}_x$, which corroborates the low photochemical activity. That is, the ambient conditions do not favor the catalytic effect of $\text{NO}_x$ to produce $\text{O}_3$. Spring is also heavily influenced by local emissions, but meteorological conditions favor photochemical activity, leading to production of the highest levels of $\text{O}_3$. In this season, winds bring background $\text{O}_3$ to the MMA, and precursors traverse the long-axis of the urban core, allowing for chemical processing, injection of fresh emissions, and high photochemical production. Spring registers the highest amount of $\text{NO}_y$, indicating that $\text{NO}_y$ is reacting efficiently with VOCs to produce photochemical oxidants, including $\text{O}_3$. Fall would seem to be influenced partly by local emissions and partly by transport. Even though fall has the second highest peak $\text{O}_3$ levels, the average $\text{O}_3$ in this season is relatively low. This can be explained by the fact that the rainy season occurs during this period. Finally, summer would be characterized by the influence of more regional (and aged) emissions. Deep planetary boundary layers are characteristic of this season, allowing mixing and dispersion of air pollutants that result in the lowest $\text{NO}_y$ levels of the year. Thus, even though temperature and solar radiation levels could suggest high $\text{O}_3$ production, photochemical activity is limited by transport and mixing effects. Finally, for the hot months (spring and summer), a distinct weekday-weekend effect can be identified, as the $\text{NO}_x$-to-$\text{NO}_y$ ratio tends to be higher during weekdays than during weekends. This would indicate changes in $\text{NO}_y$ emission rates during the week, which could lead to higher $\text{O}_3$ events during the weekend, in line with the VOC-sensitive condition of the MMA atmosphere that has been suggested by others.

Overall, the analysis of $\text{NO}_x$ and $\text{NO}_y$ levels with other chemical and meteorological variables as well as the correlation and ratios between $\text{NO}_x$ and $\text{NO}_y$ provide indicators of the level of photochemical activity that fosters $\text{O}_3$ production.
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Author details

Alberto Mendoza* and Edson R. Carrillo

*Address all correspondence to: mendoza.alberto@itesm.mx

School of Engineering and Sciences, Tecnológico de Monterrey, Monterrey, Mexico

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